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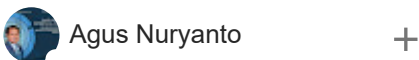
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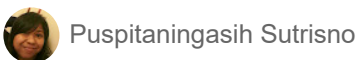
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Agus Nuryanto:

Thank you for submitting the manuscript, "Genetically continuous populations of striped snakehead (*Channa striata*) in the Cir of Biological Diversity. With the online journal management system that we are using, you will be able to track its progress thr

Submission URL: <https://smujo.id/biodiv/authorDashboard/submission/10078>

Username: agusnur

If you have any questions, please contact me. Thank you for considering this journal as a venue for your work.

Ahmad Dwi Setyawan

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Noted with thanks.

Thanks a lot.

Thank you for your response.



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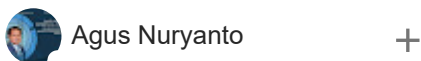
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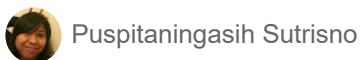
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Agus Nuryanto:

We have reached a decision regarding your submission to Biodiversitas Journal of Biological Diversity, "Genetically continuous populations of Striped Snakehead (*Channa striata*) in the Cingcingguling River fragmented by Sempor Reservoir".

Our decision is: Revisions Required

Reviewer A:

Review of "Genetically continuous populations of Striped Snakehead (*Channa striata*) in the Cingcingguling River fragmented by Sempor Reservoir".
Biological Diversity.

I have completed my throughout review of the revised manuscript (ms) and and have found the research is excellent and to be published. I recommend the manuscript for publication. I am a biologist, but also to conservation biologists of other species who may be able to utilize similar techniques in understanding the



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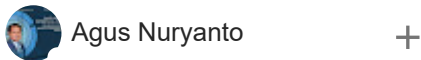
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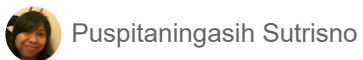
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NUNING SETYANINGRUM, W. LESTARI, KRISMONO, AGUS NURYANTO:

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Our decision is to: Accept Submission

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Agus Nuryanto:

We have reached a decision regarding your submission to Biodiversitas River fragmented by Sempor Reservoir".

Our decision is: Revisions Required

Reviewer A:

The article presents original basic research within scope of the Biodiversitas in the Cingcinggiilling River fragmented by Sempor Reservoir, Central Java

- 4. Line 236: The Sempor reservoir did not cause genetic fragmentation
- 5. Line 239: Please provide proof, data etc or supporting references

Overall, authors used clear, relatively unambiguous English language throughout

Experimental design

Research question are well defined and relevant. There is also a confusion

Genetically continuous populations of Striped Snakehead (*Channa striata*) in the Cingcingguling River fragmented by Sempor Reservoir, Central Java, Indonesia

Commented [MOU1]: Please check which one is the correct name of the reiver with the text

Abstract. Cingcingguling River, located in Kebumen Regency Central Java, Indonesia. It is fragmented by the Sempor Reservoir. The previous study proved the negative impact of the reservoir on positive rheotaxis fish, mainly in genetic constituent between the reservoir and river populations. However, research has not been conducted on the negative rheotaxis fish, such as *Channa striata*. Assessing population genetic and taxonomic validity study of Striped Snakehead in the Cingcingguling River is an essential effort. Both studies could be done using cytochrome c oxidase 1 gene. Therefore, this research aims to determine taxonomic status and evaluate the population genetic of *C. striata* in the Cingcingguling River. The samples were collected at eight sites inside and outside the reservoir. The used marker was sequenced from 53 individuals, and all specimens showed high (98.67% to 100%) and low genetic distances (0.00 to 0.01) to *C. striata* (KU692421, KU852443, and MG438366). Those values proved that all samples were genetically identified to as *Channa striata*. The vertical genetic distribution analysis proved that *C. striata* populations are genetically not different along the river. Unlike rheotaxis positive fish phenomena, the reservoir's existence does not cause genetic fragmentation and leads to continuous striped snakehead populations.

Commented [MOU2]: Which one is correct Cingcingguling or Cingcingguling? Please correct it to avoid unnecessary error.

Commented [MOU3]: Typo?

Key words: genetic diversity, reservoir, rheotaxis, striped snakehead

Abbreviations (if any): AMOVA= analysis of molecular variance, DRPM BRIN = Direktorat Riset dan Pengabdian Kepada Masyarakat Badan Riset dan Inovasi Nasional

Running title: Genetically continuous populations of *Channa striata*

INTRODUCTION

Striped snakehead (*Channa striata*) is an essential-important freshwater fish species in several Asian countries. In Indonesia, it is particularly found in the main islands of the Sunda Shelf, including Sumatra, Java, and Borneo (Froese & Pauly, 2021). Currently, this species is also found in the lesser Sunda Island, such as Bali (Yudha et al. 2018), and has been introduced to the Wallacea Regions (Irmawati et al., 2017). *C. striata* is primarily discovered in stagnant or swampy water ecosystems (Amilhat & Lorenzen 2005; Muflikhah 2007; Listyanto & Andriyanto 2009), therefore, it is a negative rheotaxis fish species. Nevertheless, it lives in a wide range of habitats, such as swamps, stagnant rivers, river flood plain, and dams or reservoirs (Iskandar & Dahiyat, 2012; Nuryanto et al., 2012; Roema, 2013; Nuryanto et al., 2015). This species is also reported found both inside and outside Sempor Reservoir in the Cingcingguling River, Central Java, Indonesia (Setyaningrum et al., 2020, 2021).

Commented [MOU4]: Is there is no previous references on this aspect?

Commented [MOU5]: Again which one is the correct one?

Commented [MOU6]: How come the fish is discovered in stagnant and swampy water and it is concluded negative? Kinda confusing readers so please explain in more detail and put it in the correct context.

Sempor Reservoir was built approximately 51 years ago and has caused the Cingcingguling River to be fragmented into two extremely different habitats. These include completely stagnant and running water bodies located underneath the reservoir (Hedianto et al., 2014). The reservoir is a physical barrier for gene flow and causes significant genetic differences among its populations (Heggnes & Roed 2006). However, the available data concerning the reservoir's negative impact on river populations was only available for the positive rheotaxis fish species (Wibowo et al., 2014; Bahiyah et al., 2013; Barasa et al. 2014; Plavova et al. 2017). Meanwhile, there is no recorded information about the reservoir's genetic effect on negative rheotaxis fish species. Therefore, it is essential to research the genetic impact of Sempor Reservoir on the *C. striata* population in the Cingcingguling River.

The genetic impact of a reservoir on the fish population could be assessed with a molecular tool, such as the cytochrome c oxidase 1 (COI) gene (Nuryanto et al. 2019). Previous research reported that it was used as a powerful robust marker for population genetic analysis of *C. striata* in Perak State situated in Malaysia (Jammaluddin et al. 2011; Tan et al. 2012, 2015). Nevertheless, population genetic research tends to be carried out when the taxonomic status of the

47 analyzed organisms is valid. In the case of *C. striata*, it was reported that the morphological identification of samples
48 obtained from different regions showed inconsistent diagnostic characters (Zhu et al. 2013; Arma et al. 2014; Khan et al.
49 2019; Muslimin et al. 2020). Taxonomic validity could also be determined using cytochrome c oxidase 1 (COI) gene (Ko
50 et al., 2013; Nuryanto et al., 2017; 2019; 2020). Furthermore, it was reliable for species delineation of *C. striata* from
51 Sumatra (Muchlisin et al. 2013; Dahruddin et al. 2016; Irmawati et al. 2017; Syaifudin et al. 2020) and is used to validate
52 morphological identification (Nuryanto et al. 2021). Therefore, this research was aimed to validate and assess the
53 taxonomic status and genetic population of *C. striata* in Sempor Reservoir Central Java using cytochrome c oxidase 1
54 gene.

55 MATERIALS AND METHODS

56 Research location and sampling sites

57 Striped snakehead specimens were collected from 8 different sites, four of them were situated inside the reservoir,
58 while the remaining were located downstream (Figure 1). These were collected using traps and lines with the help of
59 fishers. Tinny tissue samples were chopped from the pectoral fin of each specimen and preserved in ethanol 96%.
60
61
62

63
64
65 **Figure 1.** Research location with five sampling sites along the river. Four of them located inside the reservoir.
66

67 Procedures

68 Genomic DNA extraction and Marker polymerization

69 The total genome was extracted from the pectoral fin tissue using the Quick-DNA™ Miniprep Plus kit adopted from
70 Zymo's research. Extraction procedures were carried out based on the company's manual, and its success was tested using
71 1% agarose electrophoresis. Subsequently, the COI gene target fragments were reproduced using FishF2 and FishR2
72 primers (Ward et al. 2005) in Primus 25 Peqlab Thermocycler. Meanwhile, 50 µl of amplified reactions consisted of 1x
73 buffer PCR, 2 mM MgCl₂, 0.2 mM of each primer, 0.2 mM dNTP mix, 1 U Taq polymerase, and 2.0 ng / µl template
74 DNA. Furthermore, the final volume of 50 µl was adjusted by adding DNA-RNA-free water. Thermal cycles were pre-
75 denatured at 95°C for 4 minutes and were repeated 35 times. The denaturation steps lasted for 30 seconds at 95°C, 2
76 minutes at 53°C, and 1 minute at 72°C for primer annealing and chain elongation. Additionally, a final extension
77 terminated the cycles after 5 minutes, at 72°C. The PCR products were stained using ethidium bromide and 1.5% agarose
78 gel and placed under ultraviolet light. Gel documentation was further performed using the GelDoc apparatus (BioRad).

79 Marker sequencing and editing

80 The PCR products of the marker were shipped to 1st BASE Malaysia for sequencing, while that the sequencing process
 81 was performed using the Sanger method. Consensus and multiple sequences alignment were obtained by assembling the
 82 forward and reverse sequences using ClustalW ver.1.4 in Bioedit (Hall 2011). Haplotype data was obtained from its
 83 generating process in DnaSP 5 (Rozas et al. 2017).

84 Data analysis

85 The striped snakehead specimens' taxonomic status was validated through a sequence identity test carried out using a
 86 basic local alignment search tool (BLAST) closest to the taxa in GenBank. Genetic distance was also used to support the
 87 identity data. Haplotypes (h) and nucleotide (π) diversities were calculated using Arlequin 3.5, while neutral evolution of
 88 the COI marker was estimated using Fu's Fs and Tajima D test (Excoffier & Lischer 2010). Population differentiation was
 89 calculated using Fst and variance analysis (AMOVA) carried-out in Arlequin 3.5 (Excoffier & Lischer 2010). It was also
 90 estimated using a shared haplotype, which was observed in its network. This was reconstructed using the median-joining
 91 method in NETWORK software (Bandelt et al. 1999). The phylogenetic relationship of *C. striata* in Cincingguling River
 92 was estimated using Neighbor-Joining (NJ) and Maximum Likelihood algorithms in MEGA X (Kumar et al. 2018) with
 93 1000 bootstraps replications. Also, the topological stability tree was obtained from the out-group comparison (*Channa*
 94 *gacua* MK599522; *Channa micropeltes* JN024962; *Channa Lucius* KJ937433).

95 RESULTS AND DISCUSSION

96 Taxonomic status

97 Approximately 50 individuals of *Channa* specimens were successfully sequenced, resulting in fragments within the
 98 range of 596 bp to 689 bp lengths. Sequence identity test showed that the samples were genetically similar to the top 10
 99 hits closest taxa in the GenBank, all identified as *C. striata* (KU692421, KU852443, and MG438366). However, their
 100 percentages were between 98.67% and 100%, with the expected value being 0.0. The samples showed varied genetic
 101 distances in accordance with following Kimura 2 parameter (K2P) from 0.000 % to 1.019, indicating low genetic distances
 102 to their closest related taxa in GenBank, as shown in (Table 1).

103 Table 1. Sample code, expect value, percent identity, genetic distances, and closest taxa in GenBank

Sample code	E-value	Percent Identity (%)	Genetic distance (%)	Closest Taxa in GenBank
KW 1	0.0	100.0	0.000	Channa striata KU692421
KW 2	0.0	100.0	0.000	Channa striata KU692421
KW3	0.0	98.67	1.019	Channa striata KU852443
KW 4	0.0	100.0	0.000	Channa striata KU692421
KW 5	0.0	100.0	0.000	Channa striata KU692421
KW 6	0.0	100.0	0.000	Channa striata KU692421
KW 7	0.0	100.0	0.000	Channa striata KU692421
KW 8	0.0	100.0	0.000	Channa striata KU692421
KW 9	0.0	100.0	0.000	Channa striata KU692421
KW 10	0.0	100.0	0.000	Channa striata KU692421
BK 1	0.0	100.0	0.000	Channa striata KU692421
BK 2	0.0	100.0	0.000	Channa striata KU692421
BK 3	0.0	100.0	0.000	Channa striata KU692421
BK 4	0.0	100.0	0.000	Channa striata KU692421
BK 5	0.0	100.0	0.000	Channa striata KU692421
BK 6	0.0	100.0	0.000	Channa striata KU692421
KA 1	0.0	100.0	0.000	Channa striata KU692421
KA 2	0.0	100.0	0.000	Channa striata KU692421
KA 3	0.0	100.0	0.000	Channa striata KU692421
KA 4	0.0	100.0	0.000	Channa striata KU692421
KA 5	0.0	100.0	0.000	Channa striata MG438366
KA6	0.0	99.20	0.169	Channa striata KU692421
KA7	0.0	99.54	0.000	Channa striata KU692421
WO 1	0.0	99.68	0.508	Channa striata KU692421
WO 2	0.0	99.84	0.000	Channa striata KU692421
WO 3	0.0	99.35	0.848	Channa striata MG438366
WO 4	0.0	99.69	0.000	Channa striata KU692421
KS 1	0.0	99.07	0.678	Channa striata KU692421

Sample code	E-value	Percent Identity (%)	Genetic distance (%)	Closest Taxa in GenBank
KS 2	0.0	100.0	0.000	Channa striata KU692421
KS 3	0.0	100.0	0.000	Channa striata KU692421
KS 4	0.0	100.0	0.000	Channa striata KU692421
KS 5	0.0	100.0	0.000	Channa striata KU692421
KS 6	0.0	100.0	0.000	Channa striata KU692421
KS 7	0.0	98.77	0.849	Channa striata KU692421
KS 8	0.0	100.0	0.000	Channa striata KU692421
KS 9	0.0	100.0	0.000	Channa striata KU692421
KS 10	0.0	100.0	0.000	Channa striata KU692421
PW 1	0.0	99.84	0.000	Channa striata KU692421
PW 2	0.0	100.0	0.000	Channa striata KU692421
KB1	0.0	100.0	0.000	Channa striata KU692421
KB2	0.0	100.0	0.000	Channa striata KU692421
KB3	0.0	100.0	0.000	Channa striata KU692421
KB4	0.0	100.0	0.000	Channa striata KU692421
KB5	0.0	100.0	0.000	Channa striata KU692421
KB6	0.0	100.0	0.000	Channa striata KU692421
KB7	0.0	100.0	0.000	Channa striata KU692421
KB8	0.0	100.0	0.000	Channa striata KU692421
KB9	0.0	99.84	0.000	Channa striata KU692421
KB10	0.0	99.84	0.000	Channa striata KU692421
BA 1	0.0	100.0	0.000	Channa striata KU692421
BA 2	0.0	100.0	0.000	Channa striata KU692421
BA 3	0.0	100.0	0.000	Channa striata KU692421
BA 4	0.0	100.0	0.000	Channa striata KU692421

104 KW: Kedungwringin, BK: Bangkong, KA: Kalianget, WO: Waduk outlet, KS: Sempor, PW: Purbowangi, BA: Buayan, KB: Karang
105 Bolong

106 This research delineated the samples as *C. striata* because their high genetic identities (above 97%) and genetic
107 distance were less than 3%, respectively, to their conspecific. According to Ratnasingham and Hebert (2013), this is the
108 standard identity threshold for animal species determination. Simultaneously, a distance of 3% is acceptable for threshold
109 species determination in fish barcoding (Ratnasingham & Hebert 2007; Hubert et al. 2010; Candek & Kuntner 2015). Even
110 though a higher threshold of approximately 4% and 5% is allowed, other factors need to be considered (Higashi et al.,
111 2011; Jeffrey et al., 2011; Candek & Kuntner, 2015).

112 The low genetic distance among individuals of *C. striata* was reportedly occurred in the wild population found in Lake
113 Towuti, South Sulawesi, with the values between 0.043 and 0.309% and (Irmawati et al., 2017). Similar values were
114 reported in China (Zhu et al. 2013), using 5 *C. striata* populations, which showed that the intraspecific genetic distances
115 were ranged from 0.002% to 0.027%. In contrast, it was approximately 8 to 45 fold higher than among the species
116 (0.091% to 0.219%). As observed in this research, the minimum (98%) and maximum (1.019%) values of genetic identity
117 and distance, respectively, were reliably to determine the species status of the striped snakehead samples from
118 Cingcingguling Rivers. The present result is consistent with previous research carried out by Aquilino et al. (2011) and
119 Irmawati et al. (2017) that DNA barcoding is a powerful technique for species-level identification of snakehead fish.

120 Furthermore, the K2P phylogenetic tree was reconstructed by considering neighbor-joining and maximum likelihood
121 (Figure 2). Both algorithms produced a similar topology and were supported by high bootstrap values (ML=100; NJ=100).
122 *C. striata* samples formed a monophyletic clade with their conspecific reference (Figure 2). According to Xu et al. (2015)
123 and Kusbiyanto et al. (2021), monophyly is also reliable data for species determination. Figure 2 shows that the striped
124 snakehead samples and their conspecific were had a smaller branch scale than the predetermined scale of 0.02. This
125 information strongly indicates that the samples belong to the same species as their closest related taxa (*C. striata*).
126 Monophyly of *C. striata* was also detected between the natural and cultivated population in Vietnam (Nguyen & Duong
127 2015).

128 This research also indicates that the CO1 gene is a reliable marker for species identification. Its reliability serves as a
129 barcode because of this varies among species due to its high mutation rate (Sachithanandam et al., 2012). Due to its
130 variability, the CO1 gene is a suitable marker for unambiguous species identification (Balkhis et al., 2011; Winarni et al.,
131 2021). This is congruent with previous research in several locations in Indonesia (Muchlisin et al. 2013; Irmawati et al.
132 2017; Pramono et al. 2017) and other countries (Aquilino et al. 2011; Triantafyllidis et al. 2011), including Lake
133 Greece.

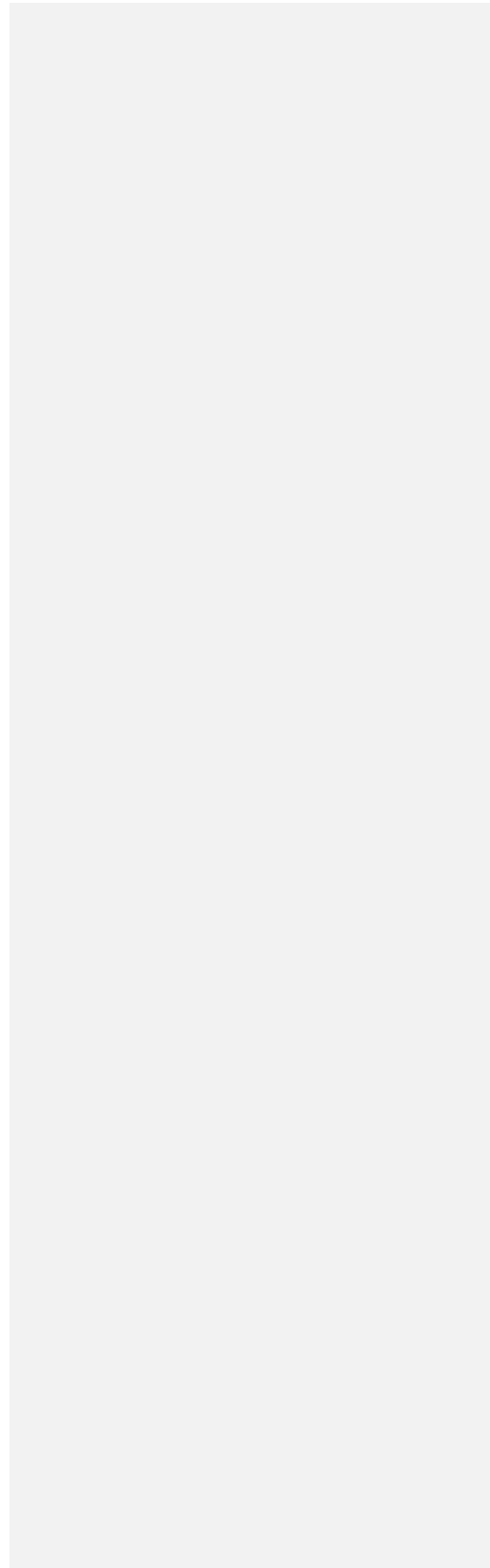
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Commented [MOU8]: But here is Ranasingham, please consistent use the correct one

Commented [MOU9]: Please specify, what This refers to?

134
135
136

Figure 2. Phylogenetic tree showing monophyly between samples and their conspecific references. left: ML value, right: NJ value



137 **Historical demography and genetic diversity**

138 Overall, Tajima's D value was -2.564, meanwhile, this significant result proved that the neutral hypothesis of marker
 139 evolution was rejected, thereby leading to the occurrence of selection pressure. However, the negative sign rejected the
 140 assumption on selection pressure and indicated a recent population bottleneck (Tajima 1989; Jong et al. 2011). The
 141 negative signs and insignificant Fu's Fs supported the neutral marker and population bottleneck assumption, as shown in
 142 Table 2. According to Jong et al. (2011) and Mohammed et al. (2021), Tajimas' D and Fu's Fs values are calculated based
 143 on haplotype and nucleotide variations, respectively. This simply signifies that Fu's Fs values are more sensitive than
 144 Tajimas' D in terms of using it for neutral theory testing of the marker.

Commented [MOU10]: Please specify this

145 **Table 2.** Population, number of individuals (N), number of haplotypes (nhp), haplotype diversity (h), nucleotide diversity (μ), Tajima' D,
 146 and Fu's Fs

Commented [MOU11]: Change to name of the Table not name of each column in the Table.

Population	N	Genetic Diversity			Neutrality Test			
		nhp	h (\pm SD?)	π (\pm SD?)	D	P	Fs	P
Overall	53	6	0.181±0.071	0.108±0.095	-2.564***	0.000	-2.360ns	0.070
SR	27	4	0.214±0.103	4.195±3.617	-2.226***	0.001	-0.913ns	0.220
KS	10	3	0.378±0.181	14.568±9.702	-1.901ns	0.006	1.726ns	0.831
PW	2	1	0.000±0.000	0.000±0.000	0.000ns	1.000	-	-
KB	10	1	0.000±0.000	0.000±0.000	0.000	1.000	-	-
BA	4	1	0.000±0.000	0.000±0.000	0.000	1.000	-	-

147 $p > 0.05 = ns$, $0.05 > p > 0.01 =$ significant, $p < 0.01 =$ highly significant, ns= non-significant, ***= highly significant

148 This research analyzed a 593 bp COI gene fragment length of 53 individual *C. striata* collected from eight sampling
 149 sites. Furthermore, it was reported that 17 out of 593 bp were polymorphic, resulting in 6 haplotypes. Overall haplotype
 150 and nucleotide diversities were 0.181 ± 0.071 , and 0.108 ± 0.095 , respectively. This indicates that *C. striata*
 151 populations in the Cingcingguling River have low genetic diversity, and this is due to 2 reasons. First, it is caused by small
 152 population sizes due to the recent bottleneck. Besides, this has been proven by negative and positive insignificant Tajimas'
 153 D and Fu's Fs values, respectively, as shown in Table 2. According to Zanella et al. (2016) and Doublet et al. (2019), a
 154 small population shows low genetic diversity due to inbreeding depression. Second, it is caused by limited ancestors, and
 155 this was proven by the haplotype network, which showed that the *C. striata* population in the Cingcingguling River had
 156 evolved from a common ancestor, as indicated in Figure 3. It was previously stated that limited maternal ancestors lead to
 157 the low genetic diversity of the offspring population because of the drift effect (Zanella et al., 2016). Besides, this attribute
 158 in *C. striata* populations was also observed in Malaysia (Jamaluddin et al., 2011).

159 The present result is inconsistent with the previous research carried out in India (Baisvar et al. 2018), stating that *C.*
 160 *striata* populations exhibited a complex pattern of genetic diversity. However, this implies that it is a common
 161 phenomenon. Meanwhile, several fish species have reported high and low haplotype genetic diversity (Sukmanomon et al.,
 162 2012; Song et al., 2013; Barasa et al., 2014; Nuryanto et al., 2020). This indicates that environmental factors have
 163 exhibited different evolutionary forces on their populations, which needs further analysis.

Commented [MOU12]: Please specify

164 The within-population evaluation indicates that the haplotype diversity of the *C. striata* population in the
 165 Cingcingguling River ranges from 0.000 ± 0.000 to 0.378 ± 0.181 . The values prove that the striped snakehead population
 166 had low genetic diversity. The majority of the populations underneath the reservoir (PW, KB, and BA) are genetically
 167 homogenous. Moreover, two subpopulations (KA and KP) had low genetic diversity. This indicates that river
 168 subpopulations show a complex genetic diversity pattern. The obtained values were lower than the previously reported
 169 results (Boonkusol & Tongbai 2016; Baisvar et al. 2018; 2019). The exploitation of *C. Striata* causes low genetic
 170 diversity, as indicated by the bottleneck effect shown by Tajimas' D and Fu's Fs values in Table 2, which caused minor
 171 population size and an opportunity for inbreeding to occur. According to Hauser et al. (2002) and Tan et al. (2012), fishing
 172 pressure reduces genetic diversity in fish species. Meanwhile, low genetic diversity caused by exploitation also occurs in
 173 various aquatic organisms and several regions (Wibowo 2012; Tan et al. 2015; Barasa et al. 2014; Baisvar. et al. 2019;
 174 Kochzius & Nuryanto 2008).

175 Table 2 shows that the nucleotide diversity ranges from 0.00 ± 0.000 to 14.568 ± 9.703 %, and these values indicate that
 176 *C. striata* in the Cingcingguling River have both low and high nucleotide diversities. According to Kochzius & Nuryanto
 177 (2008) and Nuryanto et al. (2019), when this attribute is greater than 1%, it is regarded as highly diversity. Moreover,
 178 high nucleotide diversity was detected in reservoir populations of fish species in Victoria Lake (Barasa et al., 2014).

179 **Population connectivity**

180 In accordance with Per the reservoir population, there was no difference in the genetic analysis of the four
 181 subpopulations. Therefore, this research focused on differentiating the reservoir and river populations. The amova results
 182 proved that genetic variations of -4.27% were mainly observed within the population (104.27%, Table 3), as shown in
 183 Table 3. It was assumed that no genetic differences occurred between reservoir and river populations along the
 184 Cingcingguling River, and this was supported by a negative fixation index (-0.043) and p-values of 0.115. The data proved

185 that *C. striata* populations in the Cingcingguling River formed a genetically continuous population. An interesting finding
 186 was that the Sempor Reservoir did not lead to fragmentation, as proven by the genetic similarities among the river
 187 populations. The phenomenon is related to the ecological characteristics of striped snakehead as negative rheotaxis fish,
 188 which ~~are more~~ prefer stagnant water ecosystems. The alteration of running water into a static ecosystem due to the
 189 presence of Sempor Reservoir did not significantly affect the genetics of *C. striata* both inside and outside. According to
 190 Froese and Pauly (2021), this species primarily lives in a swampy ecosystem with stagnant water.

Commented [MOU13]: Typically?

191 **Table 3.** Variance and Fst analysis among *C. striata* subpopulation

Source of variation	d.f.	Sum of square	Variance components	Percentage of variation	Fixation index (FST)	p-Value
Among subpopulations	4	0.239	-0.004 Va ^{ns}	-4.27	-0.043 ^{ns}	0.115±0.003
Within subpopulations	48	4.478	0.093 Vb	104.27		
Total	52	4.717	0.089			

193 p> 0.05 = non-significant (ns), 0.05>p>0.01= significant, p<0.01= highly significant.

194 This result is inconsistent with ~~a~~ previous research carried out by Song et al. (2013), that significant genetic structure
 195 was found among the *C. striata* population in Malaysia. However, this is due to differences in the research locations. This
 196 research examined *C. striata* in only one river but fragmented by the reservoir. In contrast, Song et al. (2013) researched
 197 different Malaysian river systems. It was previously reported that the river is a closed ecosystem and the freshwater
 198 populations tend to show significant genetic differences (Hughes 2009). Conversely, Kano et al. (2011) stated that these
 199 are solid genetic structures without physical barriers among tributaries within a river system. However, the different
 200 research locations caused an imbalance comparison between the present research and that carried out by Kano et al.
 201 (2011).

Commented [MOU14]: Please clarify about Song et al (2013). Put in the correct context of different Malaysian river systems with previously reported that the river is a closed ecosystem. It reads confusing

202 These findings were also inconsistent with the research on *Barbonymus balleroides* in the Serayu River conducted by
 203 Bahiyah et al. (2013). According to this research, the significant genetic structures between the reservoir and river
 204 population in the Serayu were observed. However, this research evaluated the population genetics of positive rheotaxis
 205 species (*B. balleroides*) whose primary habitat is running water. Therefore, the presence of reservoirs in Serayu River
 206 altered the habitat of *B. balleroides* from running to stagnant water, which became an evolutionary factor causing genetic
 207 changes in its reservoir populations. Therefore, a significant genetic structure was observed in between *B. balleroides*
 208 population inside and outside the reservoir. Furthermore, barriers, such as reservoirs, tend to cause genetic differences
 209 among the river populations (Tan et al., 2012; Adamson et al., 2012; Barasa et al., 2014). In contrast, this study observed
 210 that the Sempor Reservoir did not cause genetic fragmentation of *C. striata* because it did not alter its habitat.

Commented [MOU15]: Explain comparison in what aspect?

211 A detailed analysis of the within-population showed that the inner part of the reservoir and below the Sempor River
 212 subpopulations showed higher genetic variability than that in the downstreams (Table 2). This indicates that the upper
 213 stream subpopulations evolve faster than the lowland river regions. However, due to the age of the reservoir, which is
 214 approximately 51 years old (Hedianto et al. 2014), higher genetic variability in the upper stream subpopulations did not
 215 cause significant genetic differentiation. Similar phenomena were reported in *Chilata lopus* (Wibowo et al. 2012) and
 216 African catfish (Barasa et al. 2014).

Commented [MOU16]: Are those same or different species with Bb and Cs?

Commented [MOU17]: The Sempor reservoir did not cause genetic fragmentation of Cs because it did not alter its habitat? This sentence confuses readers. Correct it.

Commented [MOU18]: Unclear antecedent

217 Meanwhile, 6 COI haplotypes were observed in the *Channa striata* populations found in the Cingcingguling River
 218 Central Java, Indonesia. Median-joining analysis proved that haplotype 1 was dominant and found in all subpopulations,
 219 as shown in (Figure 3). The phenomenon strengthens the result of AMOVA that genetic homogeneity occurred along the
 220 Cingcingguling River, and the reservoir did not cause genetic fragmentation in the *C. striata* population.

Commented [MOU19]: Please provide proof, data etc or supporting references

221

222
223
224 **Figure 3.** Haplotype network indicates the genetic homogeneity of *C. striata* in the Cingcingguling River.

225 ● SR ● KS ● PW ● BA ● KB

226 The star-like haplotype network in (Figure 3) showed that *C. striata* populations in the Cingcingguling River evolved
227 from a single maternal ancestor (H1). The network proved that H1 is the most primitive haplotype, which is characterized
228 by high abundance and wide distribution in most populations. Similar phenomena were reported in preliminary research on
229 the *Channa* in several regions (Balkhis et al. 2011; Song et al. 2012; Adamson et al. 2012; Basvar et al. 2018; 2019) and
230 other fish groups (Barasa et al. 2014; Abila et al. 2004).

231 The snakehead fish that lives in the Cingcingguling River was genetically identified as *Channa striata* and had low
232 genetic diversity. Also, there were no genetic differences between the reservoir and river populations, which simply means
233 that *C. striata* formed a genetically homogenous population. This indicates that *C. striata* needs to be treated as a single
234 genetic conservation unit.

Commented [MOU20]: Please also relate the conclusion to genetic variability within population: Up and downstream
Line: 237-242.

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Genetically continuous populations of Striped Snakehead (*Channa striata*) in the Cingcingguling River fragmented by Sempor Reservoir, Central Java, Indonesia

Abstract. Cingcingguling River, located in Kebumen Regency Central Java, Indonesia. The Sempor Reservoir fragments it. The previous study proved the negative impact of the reservoir on positive rheotaxis fish, mainly in genetic constituents between the reservoir and river populations. However, research has not been conducted on the negative rheotaxis fish, such as *Channa striata*. Assessing population genetic and taxonomic validity study of Striped Snakehead in the Cingcingguling River is an essential effort. Both studies could be done using the cytochrome c oxidase 1 gene. Therefore, this research aims to determine taxonomic status and evaluate the population genetic of *C. striata* in the Cingcingguling River. The samples were collected at eight sites inside and outside the reservoir. The used marker was sequenced from 53 individuals, and all specimens showed high (98.67% to 100%) and low genetic distances (0.00 to 0.01) to *C. striata* (KU692421, KU852443, and MG438366). Those values proved that all samples were genetically identified as *Channa striata*. The vertical genetic distribution analysis demonstrated that *C. striata* populations are genetically not different along the river. Unlike rheotaxis positive fish phenomena, the reservoir's existence does not cause genetic fragmentation and leads to continuous striped snakehead populations.

Keywords: genetic diversity, reservoir, rheotaxis, striped snakehead

Abbreviations (if any): AMOVA= analysis of molecular variance, DRPM BRIN = Direktorat Riset dan Pengabdian Kepada Masyarakat, Badan Riset dan Inovasi Nasional

Running title: Genetically continuous populations of *Channa striata*

INTRODUCTION

Striped snakehead (*Channa striata*) is an important freshwater fish species in several Asian countries. In Indonesia, it is mainly found in the main islands of the Sunda Shelf, including Sumatra, Java, and Borneo (Adamson et al. 2010; Lakra et al. 2010; Bezinger et al. 2011; Coad et al. 2016). Currently, this species is also found in the lesser Sunda Island, such as Bali (Yudha et al. 2018), and introduced to the Wallacea Regions (Irmawati et al., 2017). *C. striata* is primarily discovered in stagnant or swampy water ecosystems (Amilhat and Lorenzen 2005; Muflikhah 2007; Listyanto and Andriyanto 2009). This species was also found both inside and outside Sempor Reservoir in the Cingcingguling River, Central Java, Indonesia (Setyaningrum et al. 2020, 2021).

Aquatic organisms are able to move in response to water currents, known as rheotaxis (Baker and Montgomery 1999; Kanter and Coombs 2006; Enders et al. 2009). Fish that actively swim against the water current is referred to as positive rheotaxis fish (Suli et al. 2012; Back-Coleman et al. 2015; Oteiza et al. 2017). In the case of *C. striata*, previous studies had reported that *C. striata* also lives in the river, but it could only be found in the parts of the river with stagnant water, river flood plain, and reservoir. It seems that *C. striata* tended to avoid water current (Iskandar and Dahiyat 2012; Nuryanto et al. 2012; Roesma 2013; Nuryanto et al. 2015). Therefore, *C. striata* could be grouped into negative rheotaxis fish. Fish species that tend to avoid water current are negative rheotaxis fish (Enders et al. 2009; Febrina 2016).

Sempor Reservoir was built approximately 51 years ago and has caused the Cingcingguling River to be fragmented into two extremely different habitats. These include entirely stagnant and running water bodies located underneath the reservoir (Hediarto et al. 2014). The reservoir is a physical barrier for gene flow and causes significant genetic differences among its populations (Heggenes and Roed 2006). However, the available data concerning the reservoir's negative impact on river populations was only available for the positive rheotaxis fish species (Wibowo et al. 2012; Bahiyah et al. 2013; Barasa et al. 2014; Plavova et al. 2017). Meanwhile, there is no recorded information about the reservoir's genetic effect on

46 negative rheotaxis fish species. Therefore, it is essential to research the genetic impact of Sempor Reservoir on the *C.*
47 *striata* population in the Cingcingguling River.

48 The genetic impact of a reservoir on the fish population could be assessed with a molecular tool, such as the
49 cytochrome c oxidase 1 (COI) gene (Nuryanto et al. 2019). Previous research reported that it was used as a robust
50 marker for population genetic analysis of *C. striata* in Perak State situated in Malaysia (Jammaluddin et al. 2011, Tan et al. 2012,
51 2015). Nevertheless, population genetic research tends to be carried out when the taxonomic status of the analyzed
52 organisms is valid. In the case of *C. striata*, it was reported that the morphological identification of samples obtained from
53 different regions showed inconsistent diagnostic characters (Zhu et al. 2013; Arma et al. 2014; Khan et al. 2019; Muslimin
54 et al. 2020). Taxonomic validity could also be determined using the cytochrome c oxidase 1 (COI) gene (Ko et al. 2013;
55 Nuryanto et al. 2017; 2019; 2020). Furthermore, it was reliable for species delineation of *C. striata* from Sumatra
56 (Muchlisin et al. 2013; Dahruddin et al. 2016; Irmawati et al. 2017; Syaifudin et al. 2020) and is used to validate
57 morphological identification (Nuryanto et al. 2021). Therefore, this research was aimed to validate and assess the
58 taxonomic status and genetic population of *C. striata* in Sempor Reservoir Central Java using cytochrome c oxidase 1
59 gene.

60

MATERIALS AND METHODS

61 Research location and sampling sites

62 Striped snakehead specimens were collected from eight different sites, four of them were situated inside the reservoir,
63 while the remaining were located downstream (Figure 1). These were collected using traps and lines with the help of
64 fishers. Tinny tissue samples were chopped from the pectoral fin of each specimen and preserved in ethanol 96%.

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70 **Figure 1.** Research location with five sampling sites along the river. Four subsampling sites are located inside the reservoir.

71

72 Procedures

73 Genomic DNA extraction and Marker polymerization

74 The total genome was extracted from the pectoral fin tissue using the Quick-DNA™ Miniprep Plus kit adopted from
75 Zymo's research. Extraction procedures were carried out based on the company's manual, and its success was tested using
76 1% agarose electrophoresis. Subsequently, the COI gene target fragments were reproduced using FishF2 and FishR2
77 primers (Ward et al. 2005) in Primus 25 Peqlab Thermocycler. Meanwhile, 50 µl of amplified reactions consisted of 1x
78 buffer PCR, 2 mM MgCl₂, 0.2 mM of each primer, 0.2 mM dNTP mix, 1 U Taq polymerase, and 2.0 ng / µl template
79 DNA. Furthermore, the final volume of 50 µl was adjusted by adding DNA-RNA-free water. Thermal cycles were pre-

80 denatured at 95°C for 4 minutes and were repeated 35 times. The denaturation steps lasted for 30 seconds at 95°C, 2
 81 minutes at 53°C, and 1 minute at 72°C for primer annealing and chain elongation. Additionally, a final extension
 82 terminated the cycles after 5 minutes, at 72°C. The PCR products were stained using ethidium bromide and 1.5% agarose
 83 gel and placed under ultraviolet light. Gel documentation was further performed using the GelDoc apparatus (BioRad).

84 *Marker sequencing and editing*

85 The PCR products of the marker were shipped to 1st BASE Malaysia for sequencing, while that the sequencing process
 86 was performed using the Sanger method. Consensus and multiple sequences alignment were obtained by assembling the
 87 forward and reverse sequences using ClustalW ver.1.4 in Bioedit (Hall 2011). Haplotype data was obtained from its
 88 generating process in DnaSP 5 (Rozas et al., 2017).

89 **Data analysis**

90 The striped snakehead specimens' taxonomic status was validated through a sequence identity test carried out using a
 91 basic local alignment search tool (BLAST) closest to the taxa in GenBank. Genetic distance was also used to support the
 92 identity data. Haplotypes (h) and nucleotide (π) diversities were calculated using Arlequin 3.5, while neutral evolution of
 93 the COI marker was estimated using Fu's Fs and Tajima D test (Excoffier and Lischer 2010). Population differentiation
 94 was calculated using Fst and variance analysis (AMOVA) in Arlequin 3.5 (Excoffier and Lischer 2010). It was also
 95 estimated using a shared haplotype observed in its network. The network was reconstructed using the median-joining
 96 method in NETWORK software (Bandelt et al. 1999). The phylogenetic relationship of *C. striata* in Cingcingguling River
 97 was estimated using Neighbor-Joining (NJ) and Maximum Likelihood algorithms in MEGA X (Kumar et al. 2018) with
 98 1000 bootstraps replications. Also, the topological stability tree was obtained from the out-group comparison (*Channa*
 99 *gacua* MK599522; *Channa micropeltes* JN024962; *Channa Lucius* KJ937433).

100

RESULTS AND DISCUSSION

101 **Taxonomic status**

102 Approximately 50 individuals of *Channa* specimens were successfully sequenced, resulting in fragments within the
 103 range of 596 bp to 689 bp lengths. Sequence identity test showed that the samples were genetically similar to the top 10
 104 hits closest taxa in the GenBank, all identified as *C. striata* (KU692421, KU852443, and MG438366). However, their
 105 percentages were between 98.67% and 100%, with the expected value being 0.0. The samples showed varied genetic
 106 distances following Kimura 2 parameter (K2P) from 0.000 % to 1.019, indicating low genetic distances to their closest
 107 related taxa in GenBank (Table 1).

108 **Table 1.** Sample code, exact value, percent identity, genetic distances, and closest taxa in GenBank

Sample code	E-value	Percent Identity (%)	Genetic distance (%)	Closest Taxa in GenBank
KW 1	0.0	100.0	0.000	Channa striata KU692421
KW 2	0.0	100.0	0.000	Channa striata KU692421
KW3	0.0	98.67	1.019	Channa striata KU852443
KW 4	0.0	100.0	0.000	Channa striata KU692421
KW 5	0.0	100.0	0.000	Channa striata KU692421
KW 6	0.0	100.0	0.000	Channa striata KU692421
KW 7	0.0	100.0	0.000	Channa striata KU692421
KW 8	0.0	100.0	0.000	Channa striata KU692421
KW 9	0.0	100.0	0.000	Channa striata KU692421
KW 10	0.0	100.0	0.000	Channa striata KU692421
BK 1	0.0	100.0	0.000	Channa striata KU692421
BK 2	0.0	100.0	0.000	Channa striata KU692421
BK 3	0.0	100.0	0.000	Channa striata KU692421
BK 4	0.0	100.0	0.000	Channa striata KU692421
BK 5	0.0	100.0	0.000	Channa striata KU692421
BK 6	0.0	100.0	0.000	Channa striata KU692421
KA 1	0.0	100.0	0.000	Channa striata KU692421
KA 2	0.0	100.0	0.000	Channa striata KU692421
KA 3	0.0	100.0	0.000	Channa striata KU692421
KA 4	0.0	100.0	0.000	Channa striata KU692421
KA 5	0.0	100.0	0.000	Channa striata MG438366
KA6	0.0	99.20	0.169	Channa striata KU692421

Sample code	E-value	Percent Identity (%)	Genetic distance (%)	Closest Taxa in GenBank
KA7	0.0	99.54	0.000	Channa striata KU692421
WO 1	0.0	99.68	0.508	Channa striata KU692421
WO 2	0.0	99.84	0.000	Channa striata KU692421
WO 3	0.0	99.35	0.848	Channa striata MG438366
WO 4	0.0	99.69	0.000	Channa striata KU692421
KS 1	0.0	99.07	0.678	Channa striata KU692421
KS 2	0.0	100.0	0.000	Channa striata KU692421
KS 3	0.0	100.0	0.000	Channa striata KU692421
KS 4	0.0	100.0	0.000	Channa striata KU692421
KS 5	0.0	100.0	0.000	Channa striata KU692421
KS 6	0.0	100.0	0.000	Channa striata KU692421
KS 7	0.0	98.77	0.849	Channa striata KU692421
KS 8	0.0	100.0	0.000	Channa striata KU692421
KS 9	0.0	100.0	0.000	Channa striata KU692421
KS 10	0.0	100.0	0.000	Channa striata KU692421
PW 1	0.0	99.84	0.000	Channa striata KU692421
PW 2	0.0	100.0	0.000	Channa striata KU692421
KB1	0.0	100.0	0.000	Channa striata KU692421
KB2	0.0	100.0	0.000	Channa striata KU692421
KB3	0.0	100.0	0.000	Channa striata KU692421
KB4	0.0	100.0	0.000	Channa striata KU692421
KB5	0.0	100.0	0.000	Channa striata KU692421
KB6	0.0	100.0	0.000	Channa striata KU692421
KB7	0.0	100.0	0.000	Channa striata KU692421
KB8	0.0	100.0	0.000	Channa striata KU692421
KB9	0.0	99.84	0.000	Channa striata KU692421
KB10	0.0	99.84	0.000	Channa striata KU692421
BA 1	0.0	100.0	0.000	Channa striata KU692421
BA 2	0.0	100.0	0.000	Channa striata KU692421
BA 3	0.0	100.0	0.000	Channa striata KU692421
BA 4	0.0	100.0	0.000	Channa striata KU692421

109 KW: Kedungwringin, BK: Bangkong, KA: Kalianget, WO: Waduk outlet, KS: Sempor, PW: Purbowangi, BA: Buayan, KB: Karang
110 Bolong

111 This research delineated the samples as *C. striata* because their high genetic identities (above 97%) and genetic
112 distance were less than 3%, respectively, to their conspecific. According to Ratnasingham and Hebert (2013), this is the
113 standard identity threshold for animal species determination. Simultaneously, a distance of 3% is acceptable for threshold
114 species determination in fish barcoding (Ratnasingham and Hebert 2007; Hubert et al. 2010; Candek and Kuntner 2015).
115 Even though a higher threshold of approximately 4% and 5% is allowed, other factors need to be considered (Higashi et al.
116 2011; Jeffrey et al. 2011; Candek and Kuntner 2015).

117 The low genetic distance among individuals of *C. striata* was reportedly occurred in the wild population found in Lake
118 Towuti, South Sulawesi, with values between 0.043 and 0.309% (Irmawati et al. 2017). Similar values were reported in
119 China (Zhu et al. 2013), using 5 *C. striata* populations, which showed that the intraspecific genetic distances were ranged
120 from 0.002% to 0.027%. In contrast, it was approximately 8 to 45 fold higher than among the species (0.091% to 0.219%).
121 As observed in this research, the minimum (98%) and maximum (1.019%) values of genetic identity and distance,
122 respectively, were reliable to determine the species status of the striped snakehead samples from Cingcingguling Rivers.
123 The present result is consistent with previous research conducted by Aquilino et al. (2011) and Irmawati et al. (2017) that
124 DNA barcoding is a powerful technique for species-level identification of snakehead fish.

125 Furthermore, the K2P phylogenetic tree was reconstructed by considering neighbor-joining and maximum likelihood
126 (Figure 2). Both algorithms produced a similar topology supported by high bootstrap values (ML=100; NJ=100). *C. striata*
127 samples formed a monophyletic clade with their conspecific reference (Figure 2). According to Xu et al. (2015) and
128 Kusbiyanto et al. (2021), monophyly is also reliable data for species determination. Figure 2 shows that the striped
129 snakehead samples and their conspecific were had a smaller branch scale than the predetermined scale of 0.02. This
130 information strongly indicates that the samples belong to the same species as their closest related taxa (*C. striata*).
131 Monophyly of *C. striata* was also detected between the natural and cultivated population in Vietnam (Nguyen and Duong
132 2015).

133 This research also indicates that the CO1 gene is a reliable marker for species identification. Its reliability serves as a
134 barcode because this gene varies among species due to its high mutation rate (Sachithanandam et al. 2012). Due to its

135 variability, the CO1 gene is a suitable marker for unambiguous species identification (Balkhis et al. 2011; Winarni et al.
136 2021). **This result** is congruent with previous research in several locations in Indonesia (Muchlisin et al. 2013; Irmawati et
137 al. 2017; Pramono et al. 2017) and other countries (Aquilino et al. al. 2011; Triantafyllidis et al. 2011), including Lake
138 Greece.

140 **Figure 2.** Phylogenetic tree showing monophyly between samples and their conspecific references. left: ML value, right: NJ value
 141

142 **Historical demography and genetic diversity**

143 Overall, Tajima's D value was -2.564; meanwhile, this significant result proved that the neutral hypothesis of marker
 144 evolution was rejected, thereby leading to selection pressure. However, the negative sign rejected the assumption on
 145 selection pressure and indicated a recent population bottleneck (Tajima 1989; Jong et al. 2011). The negative signs and
 146 insignificant F_s supported the neutral marker and population bottleneck assumption, as shown in Table 2. According
 147 to Jong et al. (2011) and Mohammed et al. (2021), Tajima's D and F_s values are calculated based on haplotype and
 148 nucleotide variations, respectively. This difference in the data used simply signifies that F_s values are more sensitive
 149 than Tajima's D in terms of using it for neutral theory testing of the marker.

150 **Table 2.** Genetic diversity value and neutrality test for the used marker

Population	N	Genetic Diversity			Neutrality Test			
		nhp	$h (\bar{x} \pm SD)$	$\pi (\bar{x} \pm SD\%)$	D	P	F_s	P
Overall	53	6	0.181±0.071	0.108±0.095	-2.564***	0.000	-2.360ns	0.070
SR	27	4	0.214±0.103	4.195±3.617	-2.226***	0.001	-0.913ns	0.220
KS	10	3	0.378±0.181	14.568±9.702	-1.901ns	0.006	1.726ns	0.831
PW	2	1	0.000±0.000	0.000±0.000	0.000ns	1.000	-	-
KB	10	1	0.000±0.000	0.000±0.000	0.000	1.000	-	-
BA	4	1	0.000±0.000	0.000±0.000	0.000	1.000	-	-

151 $p > 0.05 = ns$, $0.05 > p > 0.01 = significant$, $p < 0.01 = highly significant$, $ns = non-significant$, $*** = highly significant$

152 This research analyzed a 593 bp COI gene fragment length of 53 individual *C. striata* collected from eight sampling
 153 sites. Furthermore, it was reported that 17 out of 593 bp were polymorphic, resulting in 6 haplotypes. Overall haplotype
 154 and nucleotide diversities were 0.181 ± 0.071 , and $0.108\% \pm 0.095\%$, respectively. The data indicate that *C. striata*
 155 populations in the Cingcingguling River have low genetic diversity, and this is due to 2 reasons. First, it is caused by small
 156 population sizes due to the recent bottleneck. Besides, this has been proven by negative and positive insignificant Tajima's
 157 D and F_s values, respectively, as shown in Table 2. According to Zanella et al. (2016) and Doublet et al. (2019), a
 158 small population shows low genetic diversity due to inbreeding depression. Second, it is caused by limited ancestors. This
 159 condition was proven by the haplotype network, which showed that the *C. striata* population in the Cingcingguling River
 160 had evolved from a common ancestor, as indicated in Figure 3. It was previously stated that limited maternal ancestors
 161 lead to the low genetic diversity of the offspring population because of the drift effect (Zanella et al. 2016). Besides, this
 162 attribute in *C. striata* populations was also observed in Malaysia (Jamaluddin et al. 2011).

163 The present result is inconsistent with the previous research carried out in India (Baisvar et al. 2018), stating that *C.*
 164 *striata* populations exhibited a complex pattern of genetic diversity. However, this implies that it is a common
 165 phenomenon. Meanwhile, several fish species have reported high and low haplotype genetic diversity (Sukmanomon et al.
 166 2012; Song et al. 2013; Barasa et al. 2014; Nuryanto et al. 2020). This complex pattern of genetic diversity of *C. striata*
 167 populations indicates that environmental factors have exhibited different evolutionary forces on their populations, which
 168 needs further analysis.

169 The within-population evaluation indicates that the haplotype diversity of the *C. striata* population in the
 170 Cingcingguling River ranges from 0.000 ± 0.000 to 0.378 ± 0.181 . The values prove that the striped snakehead population
 171 had low genetic diversity. The majority of the populations underneath the reservoir (PW, KB, and BA) are genetically
 172 homogenous. Moreover, two subpopulations (KA and KP) had low genetic diversity. This data indicates that river
 173 subpopulations show a complex genetic diversity pattern. The obtained values were lower than the previously reported
 174 results (Boonkusol and Tongbai 2016; Baisvar et al. 2018; 2019). The exploitation of *C. Striata* causes low genetic
 175 diversity, as indicated by the bottleneck effect shown by Tajima's D and F_s values in Table 2, which caused minor
 176 population size and an opportunity for inbreeding to occur. According to Hauser et al. (2002) and Tan et al. (2012), fishing
 177 pressure reduces genetic diversity in fish species. Meanwhile, low genetic diversity caused by exploitation also occurs in
 178 various aquatic organisms and several regions (Kochzius and Nuryanto 2008; Wibowo 2012; Barasa et al. 2014; Tan et al.
 179 2015; Baisvar. et al. 2019).

180 Table 2 shows that the nucleotide diversity ranges from $0.00 \pm 0.000\%$ to $14.568 \pm 9.703\%$, and these values indicate that
 181 *C. striata* in the Cingcingguling River have both low and high nucleotide diversities. According to Kochzius and Nuryanto
 182 (2008) and Nuryanto et al. (2019), when this attribute is greater than 1%, it is highly diverse. Moreover, high nucleotide
 183 diversity was detected in reservoir populations of fish species in Victoria Lake (Barasa et al. 2014).

184 **Population connectivity**

185 Per the reservoir population, there was no genetic difference of the four subpopulations. Therefore, this research
 186 focused on differentiating the reservoir and river populations. The amova results demonstrated that genetic variations of -

187 4.27% were mainly observed within the population (104.27%, Table 3). It was assumed that no genetic differences
 188 occurred between reservoir and river populations along the Cingcingguling River, and this was supported by a negative
 189 fixation index (-0.043) and p-values of 0.115. The data proved that *C. striata* populations in the Cingcingguling River
 190 formed a genetically continuous population. An interesting finding was that the Sempor Reservoir did not lead to
 191 fragmentation, as proven by the genetic similarities among the river populations. The phenomenon is related to the
 192 ecological characteristics of striped snakehead as negative rheotaxis fish, which prefer stagnant water ecosystems.
 193 Alteration of running water into a static ecosystem due to Sempor Reservoir did not significantly affect the genetics of *C.*
 194 *striata* both inside and outside. Previous studies reported that *C. striata* typically lives in a swampy ecosystem with
 195 stagnant water (Amilhat and Lorenzen 2005; Muflikhah 2007; Listyanto and Andriyanto 2009).

196 **Table 3.** Variance and Fst analysis among *C. striata* subpopulation
 197

Source of variation	d.f.	Sum of square	Variance components	Percentage of variation	Fixation index (FST)	p-Value
Among subpopulations	4	0.239	-0.004 Va ^{ns}	-4.27	-0.043 ^{ns}	0.115±0.003
Within subpopulations	48	4.478	0.093 Vb	104.27		
Total	52	4.717	0.089			

198 p > 0.05 = non-significant (ns), 0.05 > p > 0.01 = significant, p < 0.01 = highly significant.

199 The present result is inconsistent with previous research carried out by Song et al. (2013), that significant genetic
 200 structure was found among rivers of the *C. striata* in Malaysia. However, this is due to differences in the research
 201 locations. This research examined *C. striata* in only one river but fragmented by the reservoir. In contrast, Song et al.
 202 (2013) researched different Malaysian river systems. It was previously reported that the river is a closed ecosystem and the
 203 freshwater populations tend to show significant genetic differences (Hughes 2009). Therefore, it was reasonable that Song
 204 et al. (2013) observed significant genetic differences among rivers. Even Kano et al. (2011) stated solid genetic structures
 205 could be observed among tributaries within a river system without physical barriers. However, the different research
 206 locations caused an imbalance comparison about genetic differentiation between the present research and that carried out
 207 by Kano et al. (2011).

208 These findings were also inconsistent with the research on *Barbonymus balleroides* in the Serayu River conducted by
 209 Bahiyah et al. (2013). According to this research, the significant genetic structures between the reservoir and river
 210 population in the Serayu were observed. However, this research evaluated the population genetics of positive rheotaxis
 211 species (*B. balleroides*) whose primary habitat is running water. Therefore, the presence of reservoirs in the Serayu River
 212 altered the habitat of *B. balleroides* from running to stagnant water, which became an evolutionary factor causing genetic
 213 changes in its reservoir populations. Therefore, a significant genetic structure was observed in between *B. balleroides*
 214 population inside and outside the reservoir.

215 It was previously reported, the presence of the Sempor Reservoir has altered upstream areas of Cingcingguling River
 216 to become flooded ecosystems or stagnant water ecosystems (Hedianto et al. 2014). Nevertheless, ecosystems alteration in
 217 the Cingcingguling River did not change the typical habitat of *C. striata*. It means that *C. striata* collected inside and
 218 outside the reservoir live in similar habitat types. Therefore, identical habitat types inside and outside the Sempor
 219 Reservoir did not cause genetic fragmentation of *C. striata*.

220 A detailed analysis of the within-population showed that WO and KS subpopulations showed slightly higher genetic
 221 variability than downstream (Table 2). This difference in genetic diversity level might indicate that both subpopulations
 222 are less exploited than the lowland river regions. The assumption arises because both subpopulations reside close to
 223 reservoirs' outlets with strong outflow. The fishers are prohibited from collecting fish near the outlet because it is
 224 hazardous (Setyaningrum et al. 2020). However, due to the age of the reservoir, which is approximately 51 years old
 225 (Hedianto et al. 2014), higher genetic variability in the upper stream subpopulations did not cause significant genetic
 226 differentiation. Similar phenomena were reported in *Chilata lopis* (Wibowo et al. 2012) and African catfish (Barasa et al.
 227 2014).

228 Meanwhile, 6 COI haplotypes were observed in the *Channa striata* populations in the Cingcingguling River Central
 229 Java, Indonesia. Median-joining analysis proved that haplotype 1 was dominant and found in all subpopulations (Figure 3).
 230 The phenomenon strengthens the result of AMOVA that genetic homogeneity occurred along the Cingcingguling River,
 231 and the reservoir did not cause genetic fragmentation in the *C. striata* population.
 232

233
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235
236

Figure 3. Haplotype network indicates the genetic homogeneity of *C. striata* in the Cingcingguling River.

● SR ● KS ● PW ● BA ● KB

237 The star-like haplotype network (Figure 3) showed that *C. striata* populations in the Cingcingguling River evolved
238 from a single maternal ancestor (H1). The network proved that H1 is the most primitive haplotype, characterized by high
239 abundance and wide distribution in most populations. Similar phenomena were reported in preliminary research on the
240 *Channa* in several regions (Balkhis et al. 2011; Song et al. 2012; Adamson et al. 2012; Basvar et al. 2018; 2019) and other
241 fish groups (Barasa et al. 2014; Abila et al. 2004).

242 The snakehead fish in the Cingcingguling River was genetically identified as *Channa striata* and had low genetic
243 diversity. The upper-stream subpopulations had higher genetic diversity than the downstream subpopulations. Also, there
244 were no genetic differences between the reservoir and river populations, which simply means that *C. striata* formed a
245 genetically homogenous population. These data indicate that *C. striata* need to be treated as a single genetic conservation
246 unit.

247

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